9 Sediment-Water Exchange

Introduction Sediment-water partitioning Particle settling and deposition Sediment erosion and resuspension Transport equation with sediments Contaminant transport within sediment bed Model of sediment-water exchange Contaminated sediment remediation

Interest in Sediments

"Geo-morphology"
Sediment as pollutant
Sediment as carrier of pollutants

Classification

Class name	Diameter (mm)	ф	Class name	Diameter (mm)	φ
Very coarse gravel	64-32	-5.5	Very fine sand	1/8-1/16	3.5
Coarse gravel	32-16	-4.5	Coarse silt	1/16-1/32	4.5
Medium gravel	16-8	-3.5	Medium silt	1/32-1/64	5.5
Fine gravel	8-4	-2.5	Fine silt	1/64-1/128	6.5
Very fine gravel	4-2	-1.5	Very fine silt	1/128-1/256	7.7
Very coarse sand	2-1	-0.5	Coarse clay	1/256-1/512	8.5
Coarse sand	1-1/2	0.5	Medium clay	1/512-1/1024	9.5
Medium sand	1/2-1/4	1.5	Fine clay	1/1024-1/2048	10.5
Fine sand	1/4-1/8	2.5	Very fine clay	1/2048-1/4096	11.5

ASCE, 1975; φ= -ln(d_{mm})/ln(2)

Most environmental interest in finer fractions



Equilibrium Partitioning

Partition (or distribution) coefficient:

 $K_{p} = \frac{C_{s}}{C_{d}}$ Simplest model:

two phases

Sorbed phase conc c_s (mass contaminant per mass solid) divided by dissolved phase conc c_d (mass contaminant per volume solvent) in equilibrium

Units of K_p: volume/mass, e.g., cm³/g

More complicated partitioning models

 K_p depends on contaminant, its concentration, concentration of organic matter, redox potential, etc. Typical values: 10¹ to 10² cm³/g (hydrophilic) to 10⁴ to 10⁵ cm³/g (hydrophobic)

In sediment bed

Mass of dissolved contaminant/volume



Mass of sorbed contaminant/volume

Ratio of sorbed to total mass

 $=\phi c_d$

$$= \rho_s c_s (1 - \phi) = \rho_s K_p c_d (1 - \phi)$$

if equilibrium

Unit volume with porosity ϕ

$$f = \frac{\rho_{s} K_{p} (1 - \phi)}{\rho_{s} K_{p} (1 - \phi) + \phi} = \frac{K_{p} \rho}{K_{p} \rho + 1}$$

$$\rho = \rho_s (1 - \phi) / \phi = \rho_{b,d} / \phi$$

$$\rho_{b,d} = \rho_s (1 - \phi)$$

$$\rho_{b,w} = \rho_s (1 - \phi) + \phi$$

Solid-water phase ratio (solid mass/water mass) Bulk (dry) sediment density

Bulk (wet) sediment density

Surficial sediments

 $\rho_{\rm s} \sim 1.5 - 2.5 \text{ g/cm}^3; \phi \sim 0.6 - 0.8 =>$

 $\rho_{b,d} \sim 0.3-1 \text{ g/cm}^3$, $\rho_{b,w} \sim 1.1-1.6 \text{ g/cm}^3$, $\rho \sim 0.4-1.7 \text{ g/cm}^3$



While most of the contaminant is associated with solids, the dissolved phase is very important because it is more bio-available and amenable to sediment-water exchange

Sediment quality criteria often derived from target dissolved phase concentrations assuming equilibrium partitioning

In water column

"porosity" ~ 1 so $\rho = \rho_s (1 - \phi) / \phi = [TSS]$

[TSS] ~ 1 to 100 mg/L (10⁻⁴ to 10⁻⁶ g/cm³)

$$F = \frac{K_p \rho}{K_p \rho + 1} = \frac{K_p [TSS]}{K_p [TSS] + 1} = \frac{(10^{-4} \text{ to } 10^{-6})K_p}{(10^{-4} \text{ to } 10^{-6})K_p + 1}$$

For hydrophobic contaminants ($K_p \sim 10^4$ to 10^5) concentrations in sorbed and dissolved phases can be comparable. Hydrophylic contaminants are mostly in dissolved phase

Non-equilibrium conditions



K Rate constant (t^{-1}); e.g. (Wu and Gschwend, 1986)



Non-equilibrium, cont'd

 κ/K_{p}

Time scale for desorption from particle = days to months => equilibrium assumption not very good for suspended sediments (may be OK for stationary sediments)

 $\frac{d}{dt}(c_d + \rho c_s) = 0$ Total mass is conserved

Additional Comments

Fine particles usually most important

- Most easily resuspended
- Settle most slowly
- Probably have highest f_{oc} => K_p
- Highest $\kappa \sim (diameter)^{-2}$
- Models often have multiple particle sizes with individual settling velocity, tendency to resuspend, and κ

Sediment movement



Modes of transport

Settling & deposition (non-cohesive and cohesive)
 Erosion & resuspension (non-cohesive and cohesive)
 Bed-load (non-cohesive)

Particle settling (WWT jargon)

Discrete (Type 1) (non-cohesive)
Flocculant (Type 2) (cohesive)
Hindered or zone (Type 3)
Compression (Type 4)

Focus on Types 1 and 2

Terminal velocity



 $C_d = f(R); R = Reynolds number = w_s d/v$

Applicable to cohesive and non-cohesive sediments, but aggregate shapes and densities are ill-defined for cohesive sediments

Spherical particles

R < 1; $d < 100 \mu m$ (Stokes settling)

$$F_{d} = 3\pi\rho\upsilon dw_{s}$$

$$C_{d} = \frac{24}{R} \qquad w_{s} = \frac{g(s-1)d^{2}}{18\upsilon}$$

 $R < 10^4$ (Metcalf & Eddy, 1991)

$$C_d = \frac{24}{R} + \frac{3}{\sqrt{R}} + 0.34$$

More generally (0.01 mm < d < 100 mm; Dietrich, 1982)

 $\log w^* = -3.76715 + 1.92944(\log D^*) - 0.09815(\log D^{*2})$

 $-0.00557(\log D^{*3}) + 0.00056(\log D^{*4})$

$$w^* = \frac{w_s^3}{(s-1)g\nu} \qquad D^* = \frac{(s-1)gd^3}{\nu^2}$$

Spherical & non-spherical particles



Settling of Cohesive Sediments



Figure by MIT OCW.

Settling velocity depends on concentration. Empirical formula from EPA (2004):

$$w_{s}(m/d) = \frac{0.1m/dC_{WL} + 30m/d(C_{COH} - C_{WL})}{C_{COH}}$$

 C_{WL} = wash load concentration (5 mg/L); C_{COH} = cohesive sediment concentration (mg/L). Within range of empirical observations, e.g., Hawley (1982)

Sewage particles

Table 9.2 Summary of settling velocities (percent settling faster than indicated value)(after Stolzenbach and Adams, 1998)

W _s (cm/sec)	EPA (Pr Effl	(1982) im Raw	Wang Holo Effl ¹	g (1988) ography Sludge ¹	Wang (1988) Column Sludge ^{2,4}	Faisst (1976, 1980) Sludge	Ozturgut & Lavelle (1986) Effl ³	Lavelle et al. (1988) Sludge ⁴	McCave & Gross (1991) Nat sed	Albro et al. (1996) Sludge
1		5				_	_	·····		4-15
10-1	5	40	_		0-25	0-5	_	10-20	-	26-55
10 ⁻²	20	60	≈3	≈3	10-60	5-40	5-25	25-30	10	>55 - 84 ⁵
10 ⁻³	50	85	≈10	≈30	30-75	25-70	50-60	50-80	40-45	_
10-4	_		≈20	≈60	45-85	50-80	≈70	95-100	100	

¹Holography

²Column

³Computed

⁴Range for sludge reflects coagulation; highly coagulated samples settle faster

⁵Indicated percentages are for 0.04 cm/sec

Often a wide range of settling velocities

Settling Basins: discrete settling



Inlet

Outlet

Time to settle $T_s = h/w_s$ Hydraulic residence time $T_{res} = V/Q = Ah/Q$ $T_{res} > T_s => Ahw_{sc}/Qh > 1$ $w_{sc} > Q/A$ or $Q < Aw_{sc}$ or $A > Q/w_{sc}$

Comments

- Q/A = overflow rate (really a velocity); particles settling faster will deposit
- Flow capacity depends on area (not depth) => make tanks as shallow as practical
- Increase area using stacked clarifiers, inclined plates/tubes, etc.
- Similar concepts apply to field: deposition at river deltas, in ponds & reservoirs, downstream from outfalls
- Example of particulate phosphorus loading in reservoir (Vollenweider plots)

Retention in lakes & reservoirs: critical phosphorus loading



Phosphorous loading diagram



After Vollenweider (1975); Chapra (1997)

Particle scavenging

Removal of marine contaminants (e.g., metals) by natural particle settling





Particle scavenging, cont'd

With deposition, ²³⁴Th conc c < c_e

 $f = fraction of ^{234}Th sorbed to particles$



Particle scavenging, cont'd

Use ²³⁴Th deposition to trace deposition of another metal (call it x), with different partitioning



Deposition

Flux of settling particles $w_s C$ C = sediment concentration (TSS) Deposition rate, D

 $D = pw_s C$ p = probability of depositing (sticking)

For non-cohesive sediments, p = 1

For cohesive sediments

$$p = (1 - \tau / \tau_{c,d})$$
 $\tau < \tau_{c,d}$ $\tau_{c,d} = critical depositional shear stress$

 $0.06 < \tau_{c,d} < 1.1 \text{ N/m}^2$ (Mehta & Partheniades, 1975; Ziegler et al, 1995)

Particle accumulation on bottom



How to determine $w_o = dh/dt$?

- Depth of natural, accidental or intentional marker
 - e.g., paint pigments in Fort Point Channel
- Decay of radioactive tracer
 - e.g., ²¹⁰Pb,

Fort Point Channel

Recall discussion in Chapter 4



Fluorescent dye and pigment particles released May '90 and July '91

Figure by MIT OCW.



Stolzenbach and Adams (1998)

Pigment surveyed with freeze corer



Comments

Deposition rates of 1-4 cm/yr in FPC (three cores)
 Loading rates for all FPC sediment sources ~ 0.14 cm/yr
 Substantial import of (contaminated) sediment



Measuring deposition with Excess ²¹⁰Pb



Pb-210 particle reactive tracer; $t_{1/2} \sim 23$ yr c = excess concentration Relative to (moving) interface, steady state, no sediment mixing

$$-w_{o}\frac{dc}{dz} = -\lambda c$$
$$\frac{c}{c_{o}} = e^{-\frac{\lambda z}{w_{o}}}$$
Alternatively

$$w_o = \frac{1}{c_o} \lambda \int_{-\infty} c(z) dz$$

Erosion and Resuspension

Associated with bottom shear stress


Erosion of non-cohesive seds

 $\$ Critical shear stress τ_{c} required to initiate particle motion



 $(\tau/\rho)^{0.5}$ > w_s, τ_c => suspended load

Re-suspension flux depends on near-bed concentration; many formulations



Cohesive sediments

Many formulations; most apply for shear stress above a critical erosional shear stress, $\tau_{c,e}$

Erosion rate E (g/m²-s)

$$E = M \left(\frac{\tau}{\tau_{c,e}} - 1\right)^n$$

 $\tau_{c,e} = 0.05 \text{ to } 0.3 \text{ N/m}^2;$ M = 0.1 to 3 g/m²-s; n = 1-3

Cohesive sediments, cont'd

Erosion potential ϵ (g/m²)

$$\varepsilon = \frac{a_o}{T_d^m} (\frac{\tau}{\tau_{c,e}} - 1)^n$$
 $T_d = time (days) after deposition$

 $\tau_{c,e} = 0.1 \text{ N/m}^2$; $a_o = 50$; m = 2; n = 2.7

Erosion over specified time interval ~ 1 hr

Above parameters from Ziegler, et al., 1995 for Watts Bar Reservoir, TN

Measurement of erosion

 Linear laboratory flume
Linear flume in field
Laboratory annular flume
Portable resuspension device (Shaker; Tsai and Lick, 1986)



Measures erosion rates in the lab



Comments

Regions with $\tau > \tau_{c,e}$ on regular or intermittent basis exhibit erosional tendencies

Transport Equation with Sediments

$$\frac{\partial c}{\partial t} + \frac{\partial}{\partial x}(uc) + \frac{\partial}{\partial y}(vc) + \frac{\partial}{\partial z}(w - w_s)c = \frac{\partial}{\partial x}(E_x\frac{\partial c}{\partial x}) + \frac{\partial}{\partial y}(E_y\frac{\partial c}{\partial y}) + \frac{\partial}{\partial z}(E_z\frac{\partial c}{\partial z})$$

z positive upward, origin at sediment interface; c = sediment concentration; w_s = settling velocity

Neglecting horizontal transport & vertical water velocity

$$\frac{\partial c}{\partial t} - \frac{\partial}{\partial z} w_s c = \frac{\partial}{\partial z} (E_z \frac{\partial c}{\partial z})$$

Boundary conditions

$$-w_{s}c - E_{z}\frac{\partial c}{\partial z} = 0 \quad \text{at surface } (z = h)$$
$$-w_{s}c - E_{z}\frac{\partial c}{\partial z} = (\alpha - 1)w_{s}c \quad \text{at bottom } (z = 0)$$

Surface and bottom BCs



 α denotes relative amount of erosion

Vertical sediment distribution

Under steady state

$$-w_s c = E_z \frac{dc}{dz}$$

Logarithmic velocity profile in channel; turbulent diffusivity = viscosity

$$\frac{c(z)}{c_a} = \left[\left(\frac{h-z}{z} \right) \left(\frac{a}{h-a} \right) \right]^{w_s/\kappa u}$$

 $w_s / \kappa u_*$ Rouse number

Depth-average E_z (= 0.07u_{*}h)

 $Pe = \frac{w_s h}{E_z}$ Peclet number ~ 6 w_s/ κ u_{*}





Dhamothran et al (1981)

Figure by MIT OCW.



 $w_s = 10^{-2}$ to 10^{-6} m/s (Table 9.2)

Table 9.2 Summary of settling velocities (percent settling faster than indicated value) (after Stolzenbach and Adams, 1998)

w _s (cm/sec)	EPA (Pr Effl	(1982) im Raw	Wang Holo Effl ¹	g (1988) ography Sludge ¹	Wang (1988) Column Sludge ^{2,4}	Faisst (1976, 1980) Sludge	Ozturgut & Lavelle (1986) Effl ³	Lavelle et al. (1988) Sludge ⁴	McCave & Gross (1991) Nat sed	Albro et al. (1996) Sludge
1		5		_	*****	_	_			4-15
10-1	5	40	_		0-25	0-5	_	10-20	_	26-55
10-2	20	60	≈3	≈3	10-60	5-40	5-25	25-30	10	>55-84 ⁵
10 ⁻³	50	85	≈10	≈30	30-75	25-70	50-60	50-80	40-45	_
10-4	-		≈20	≈60	45-85	50-80	≈70	95-100	100	
¹ Holograp	phy									

²Column

³Computed

⁴Range for sludge reflects coagulation; highly coagulated samples settle faster ⁵Indicated percentages are for 0.04 cm/sec

Often a wide range of settling velocities



Pe for settling basin and river

W _s	$Pe = w_sh/E_z$	$Pe = w_sh/E_z$
(m/s)	(Basin)	(River)
10-2	340	
10 -3	34	
10-4	3.4	0.1
10-5	0.34	0.01
10-6	0.034	0.001



 $Pe = \sim 300 w_s/u$ (third column of Table 9.3)

Comments

In basin, turbulence insufficient to mix particles that settle (Pe > 30) In river, turbulence sufficient to mix particles that don't settle in basin (Pe <0.1) (river can be treated as well mixed) •In basin, $\tau_{b} = \rho u_{*}^{2} = 0.07 \text{ N/m}^{2} < \tau_{c.e}$ •In river, $\tau_{b} = \rho u_{*}^{2} = 0.22 \text{ N/m}^{2} < \tau_{c.e}$ (possible resuspension)

Vertically well-mixed conditions

Pe < 0.2

3-D equation

$$\frac{\partial c}{\partial t} - \frac{\partial}{\partial z} w_s c = \frac{\partial}{\partial z} (E_z \frac{\partial c}{\partial z})$$

Vertical integration

$$h\frac{\partial \overline{c}}{\partial t} = \left[w_s c + E_z \frac{\partial \overline{c}}{\partial z}\right]_{surf} - \left[w_s c + E \frac{\partial \overline{c}}{\partial z}\right]_{bot}$$
$$= 0$$

Vertically well-mixed conditions, cont'd

No resuspension ($\alpha = 0$)

$$\frac{d\overline{c}}{dt} = -\frac{w_s\overline{c}}{h}$$

 $\overline{c} = c_o \exp(-w_s t/h)$

 $c_o = initial depth-averaged concentration$

 $w_s/h = first order removal rate, \kappa_s$

Partially-mixed conditions sometimes analyzed using $\kappa_s > w_s/h$ (because near bottom concentrations are greater than c_o)

Multiple size fractions



Contaminant transport within & across the sediment bed



Porewater advection (GW movement; sediment compaction; wave or bedform induced pressures; biomixing

$$J_a = \phi u c_d$$

Contaminant transport within & across the sediment bed



Porewater advection (GW movement; sediment compaction; wave or bedform induced pressures; biomixing

 $\boldsymbol{J}_a = \boldsymbol{\phi} \boldsymbol{\mu} \boldsymbol{c}_d$

Porewater diffusion

$$J_d = -\phi D' dc_d / dz$$

$$D' = \phi D_m$$

Contaminant transport within & across the sediment bed



Porewater advection (GW movement; sediment compaction; wave or bedform induced pressures; biomixing

Bulk sediment motion ("turbulence")

$$J_b = -D_b d(c_d + \rho_s c_s) / dz$$

Porewater diffusion

$$J_d = -\phi D' dc_d / dz$$

 $J_a = \phi u c_d$

$$D' = \phi D_m$$

Sediment Profile Imaging



 Benthic fauna mix dissolved oxygen and other sediment characteristics (oxygen rich areas are light colored)
Note feeding tubes near surface

EPA, 2006

Measuring bioturbation with ²³⁴Th



Th-234 particle reactive tracer (c); $t_{1/2} \sim 24.1$ day

Relative to (moving) interface, steady state, including sediment mixing

$$-w_o \frac{dc}{dz} = D_b \frac{d^2 c_s}{dz^2} - \lambda c$$

For $D_b \lambda >> w_o$

$$0 = D_b \frac{d^2 c_s}{dz^2} - \lambda c$$

$$\frac{c}{c_o} = \exp(-\sqrt{\lambda/D_b}z)$$

Comments

D_b correlates with w_o (reflecting flux of organic matter)
Coastal sediments: D_b = 10⁻⁷ to 10⁻⁶ cm²/s
Deep sea sediments: 10⁻⁹ to 10⁻⁸ cm²/s

DDT on Palos Verdes Shelf (WE 9-4)



DDT commonly used pesticide until 1970s (Silent Spring). $\diamond \sim 1700$ T discharged by LACSD's White Point outfall (60m depth) (also agricultural run-off) ♦ ~100T (p-p'-DDE) still buried in sediment Issues of environmental racism EPA Superfund Site (Montrose Chemical Co.)

Vertical Profiles

Core 8C 1981 (solid) to 1989 (open)



Vertical distribution of porosity (open squares) and bioturbation coefficient (closed squares)



Exponential distribution of porosity and bioturbation (latter based on worm density)

Concentration vs depth

(USGS; Lee, 1994)

Issues

Contamination slowly decreasing. But is it biodegradation or surface loss? Will natural sedimentation cap contaminants? Decreasing since WWTP upgrade; introduce clean sediments from flood control reservoirs? Current strategy of institutional controls (public) outreach, fish monitoring, etc.) Is this enough? Possible future capping. Will this work? (2000) pilot capping failed.)

Sediment Fate Processes



Deposition of clean sediment (deposition velocity w in cm/yr) Biological mixing (D_h) in cm²/yr) Biodegradation (1st O) rate λ in yr⁻¹) Release to surface (k) in cm/yr)

$$J = k\rho_s (1 - \phi)c_{so}$$

Mass Transport in Sediments

$$(1-\phi)\rho_{s}\frac{\partial c_{s}}{\partial t} + \frac{\partial}{\partial \varsigma}[w(1-\phi)\rho_{s}c_{s}] = \rho_{s}\frac{\partial}{\partial \varsigma}\left\{D_{b}\frac{\partial}{\partial \varsigma}[(1-\phi)c_{s}]\right\} - \lambda(1-\phi)\rho_{s}c_{s}$$

Boundary conditions

$$D_{b} \frac{\partial c_{s}}{\partial \varsigma} = (k + w_{o})c_{s} \text{ at } \varsigma = 0 \qquad \zeta = \text{depth below} \\ (\text{moving}) \text{ sediment bed} \\ c_{s} = 0 \qquad \text{at } \varsigma = \infty$$

Use observations to calibrate unknown parameters w, D_b , λ and k

Simplification



Spatial Moments



Sediment water exchange model



Figure by MIT OCW.

 Steady state
Includes bioturbation, pore-water diffusion and sorption kinetics, but no resuspension, deposition or bio-degradation
Colloidal transport included but not described here

 Applied to PAH's in Boston Harbor

Chen, 1993

Sediment water exchange model



Boundary Conditions

$$c_d = c_{d1}$$
 $dc_s/dz = 0$ at $z = 0$

$$c_d K_p = c_s = c_{sL}$$
 at $z = L$
Approximate Solution



Flux to surface Dissolved phase concentration in equilibrium with c_{sL} $J = \frac{c_{sL}/K_p}{\frac{\tau}{H} + \frac{z_w}{D_m} + \frac{2.2R}{(1-\phi)[(D_b + D')D_m\rho_s K_p]^{1/2}} + \frac{L}{(1-\phi)\rho_s D_b K_p}}$ 1 2 3 4

Denominator: 4 "resisters" in series: 1) flushing, 2) water-side diffusion, 3) sorption kinetics, 4) bio-mixing

Parameters

Varia ble	Definition	Value(s)
D _m	Aqueous solution diffusivity	0.8x10 ⁻⁵ cm²/s
D'	Aqueous solution diffusivity corrected for porosity	0.5x10 ⁻⁵
D _b	Bioturbation coefficient	10 ⁻⁷ , 10 ⁻⁶ , 10 ⁻⁵ cm ² /s
K _p	Solid-water partition coefficient	10 ¹ to 10 ⁶ cm ³ /g
L	Biologically active depth	5 cm
φ	Porosity	0.8
ρ _s	Sediment density	2.5 g/cm ³
	Sorbed concentration at z = L	10 ⁻⁶ g/g
R	Characteristic aggregate radius	0.01 cm
	Water-side boundary layer thickness	0.06 cm
	Hydrodynamic residence time of overlying water	5 day
Н	Depth of overlying waterbody	6 m
	Desorption rate constant	Eq (9.5)







Comments

Water side bl (2) controls for large K_p & D_b
Bioturbation (4) controls for small K_p & D_b

(Resistance on side with smallest equilibrium concentration)

Desorption not limiting factor

Longest clean-up times for high K_p (nearly a century for benzo(a) pyrene (K_p ~ 10⁵) in Boston Harbor)

Dealing with Contaminated Sediment

Natural attenuation (Let it sit)

- If evidence of natural recovery (deposition, biodegradation)
- Or if other options problematic
- Combined w/ active monitoring & inst controls

Capping (Cover it up)

- With clean sediment
- In situ or in confined aquatic disposal (CAD) cells
- Dredging (Remove it)
 - Environmental (remove contamination)
 - Maintenance (keep harbors/channels open)
 - Improvement (make harbors/channels deeper)

Boston Harbor Navigation Improvement Project





 Deepen to 38-40' (versus maintenance or environmental)
1.7x10⁶ yd³ clay (MBDS)
1.1x10⁶ yd³ silt (CAD cells)

Confined Aquatic Disposal Cells





Figure by MIT OCW.

Dredge buckets



Environmental Clam shell

CAD Challenges





Subbottom line 6-003 from cell M4 (OSI 1999), annotated at bottom showing location of cores, fluidized mud layer (above red dashed line), sand zone (between red and blue dashed lines), and approximate bottom of cell (green dashed line). Note reversal of East and West. Cell M4 - Post-Cap Sub-Bottom Profile



Verifying CAP integrity

Waiting for sufficient consolidation

Additional Issues

 Containing dredged and capping material (during descent & upon impact)
Time of disposal (environmental windows to allow migrating fish passage)
Residual silt (should you "rake all the leaves?")

Open cells (exposure to uncapped material)