LECTURE 13

MONITORED NATURAL ATTENUATION

Monitored Natural Attenuation

As defined by U.S. EPA:

reliance on natural processes to achieve site-specific remedial objectives

Source: Pope, D. F., and J. N. Jones, 1999. Monitored Natural Attenuation of Chlorinated Solvents. U.S. EPA Remedial Technology Fact Sheet. Report Number EPA/600/F-98/022. Office of Research and Development, U.S. Environmental Protection Agency, Washington, D.C. May 1999. (http://www.epa.gov/ada/download/fact/chl-solv.pdf).



Historical development of MNA

Historical development

- 1985 Always a remedial alternative: EPA says it was used in Superfund as early as 1985
- 1988 Guidance on Remedial Actions for Contaminated Ground Water at Superfund Sites discusses natural attenuation, but mostly as a comparison standard for active remediation. Natural attenuation is not encouraged
- Began to be more commonplace with recognition of intractability of DNAPL cleanups and inadequacy of pump and treat technology
- Simultaneously, there was increasing recognition that in situ processes were containing or cleaning up contamination
- 1993-94 Two key studies by National Research Council:
 - In Situ Bioremediation When does it work? 1993 study discusses intrinsic bioremediation
 - Alternatives for Ground Water Cleanup 1994
- 1995 Technical Protocol for Implementing Intrinsic Remediation with Long-Term Monitoring for Natural Attenuation of Fuel Contamination Dissolved in Groundwater
 - Prepared by Air Force Center for Environmental Excellence working with U.S. EPA research laboratory
 - Defines procedure to show intrinsic remediation is occurring
- September 1996 EPA Symposium on Natural Attenuation of Chlorinated Organics in Ground Water
- November 1996 Draft Technical Protocol for Evaluating Natural Attenuation of Chlorinated Solvents in Groundwater (Final in Sept. 1998)
- December 1997 OSWER Policy Directive: Use of Monitored Natural Attenuation at Superfund, RCRA Corrective Action, and Underground Storage Tank Sites
- Sept.-Dec. 1998 EPA Seminars on Monitored Natural Attenuation in nine cities around the US

1988 Guidance on Ground Water

United States Environmental Protection Agency

Office of Emergency and Remedial Response Washington DC 20460 EPA/540/G-88/003 OSWER Directive 9283.1-2 December 1988

Superfund



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Guidance on Remedial Actions for Contaminated Ground Water at Superfund Sites

1993 and 1994 – NRC Studies

"In Situ Bioremediation. When does it work?" National Research Council.

"Alternatives for Ground Water Cleanup." National Research Council.

1995 – AFCEE Technical Protocol

Technical Protocol for Implementing Intrinsic Remediation with Long-Term Monitoring for Natural Attenuation of Fuel Contamination Dissolved in Groundwater

Volume 1

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1996 – EPA Symposium on Natural Attenuation







United States Environmental Protection Agency Office of Research and Development Washington, DC 20460 EPA/540/R-96/509 September 1996

Symposium on Natural Attenuation of Chlorinated Organics in Ground Water

Hyatt Regency Dallas Dallas, TX September 11–13, 1996



1996 – Draft Technical Protocol

United States Environmental Protection Agency Office of Research and Development Washington DC 20460 EPA/600/R-98/128 September 1998

\$EPA

Technical Protocol for Evaluating Natural Attenuation of Chlorinated Solvents in Ground Water

Nov. 1996 – Draft Sept. 1998 - Final



1997 – MNA Policy Directive

United States Environmental Protection Agency Office of Solid Waste and Emergency Response

€EPA DIRECTIVE NUMBER: 9200.4-17 TITLE: Use of Monitored Natural Attenuation at Superfund, RCRA Corrective Action, and Underground Storage Tank Sites APPROVAL DATE: December 1, 1997 EFFECTIVE DATE: December 1, 1997 **ORIGINATING OFFICE: OSWER** FINAL X **DRAFT** Interim Final STATUS: **REFERENCE** (other documents):

OSWER OSWER OSWER DIRECTIVE DIRECTIVE DIRECTIVE

1998 – MNA Seminars

SEDA

United States Environmental Protection Agency Office of Research and Development Washington, DC 20460 EPA/625/K-98/001 September 1998

Seminars

Monitored Natural Attenuation for Ground Water

September 2–3, 1998—Philadelphia, PA September 14–15, 1998—Denver, CO September 16–17, 1998—Chicago, IL October 14–15, 1998—Kansas City, MO November 2–3, 1998—Dallas, TX November 16-17, 1998-Atlanta, GA December 2–3, 1998—Seattle, WA December 8–9, 1998—Boston, MA December 14–15, 1998—San Francisco, CA

Definition of MNA (continued)

May be physical, chemical or biological Act without human intervention to reduce the mass, toxicity, mobility, volume, or concentration of contaminants in soil or ground water

- Processes include:
 - biodegradation
 - dispersion
 - dilution
 - sorption
 - volatilization

chemical or biological stabilization, transformation, or destruction

Source: Pope, D. F., and J. N. Jones, 1999. Monitored Natural Attenuation of Chlorinated Solvents. U.S. EPA Remedial Technology Fact Sheet. Report Number EPA/600/F-98/022. Office of Research and Development, U.S. Environmental Protection Agency, Washington, D.C. May 1999. (http://www.epa.gov/ada/download/fact/chl-solv.pdf).

Elements of Natural Attenuation

MNA is not:

- No-action alternative
- Presumptive or default remedy

MNA must be:

- Evaluated along with other alternatives
- Selected only is it meets remediation objectives
- Work in reasonable time frame (30 years)
- Used very cautiously as sole remedy

Components of MNA

Required components of MNA:

- Source control
- Performance monitoring

Prerequisite for MNA:

Site-specific characterization data and analysis

Demonstrating efficacy of MNA

- 1. Historical chemical data showing clear trend of decreasing mass or concentration.
- 2. Hydrogeologic or geochemical data that indirectly demonstrate natural attenuation processes
- 3. Field or microcosm studies that directly demonstrate natural attenuation processes

Aerobic biodegradation of fuel hydrocarbons

Oxygen used as electron acceptor

 $2\mathrm{C}_{6}\mathrm{H}_{6} + 15\mathrm{O}_{2} \rightarrow 12\mathrm{CO}_{2} + 6\mathrm{H}_{2}\mathrm{O}$

Indicators of biodegradation:

Reduction in dissolved oxygen (3 mg DO needed to metabolize 1 mg of benzene)

Reduction in hydrocarbon concentration

Biodegradation sequence

Order of aerobic biodegradation:

- Ethyl benzene
- Toluene
- Benzene
- **Xylene**

Denitrification

Nitrate is electron acceptor

$$6NO_3^- + 6H^+ + C_6H_6 \rightarrow 6CO_2^\uparrow + 6H_2O + 3N_2^\uparrow$$

Actually occurs in multiple steps mediated by different bacteria:

$$NO_3^- \rightarrow NO_2^- \rightarrow NO \rightarrow N_2O \rightarrow NH_4^+ \rightarrow N_2$$

Denitrification

Indicators of biodegradation:

- Reduction in nitrate
- Reduction in hydrocarbon concentration
- Presence of denitrifying bacteria
- Reducing conditions (dissolved oxygen < 1 mg/L)

Iron reduction

Insoluble iron(III) (ferric iron) is electron acceptor Reduced to soluble iron(II) (ferrous iron)

 $\begin{array}{r} 60 \mathrm{H^{\scriptscriptstyle +}} + 30 \mathrm{Fe}(\mathrm{OH})_3 + \mathrm{C_6H_6} \rightarrow \\ & 6 \mathrm{CO_2} + 30 \mathrm{Fe^{2+}} + 78 \mathrm{H_2O} \end{array}$

Iron reduction

Indicators of iron biodegradation: Increase in dissolved iron Reduction in hydrocarbon concentration Low or no dissolved oxygen

Sulfate reduction

Sulfate is electron acceptor Sulfate reduction to sulfide

 $30H^{+} + 15SO_{4}^{2-} + 4C_{6}H_{6} \rightarrow 24CO_{2} + 15H_{2}S + 12H_{2}O$

Methanogenesis (Methane fermentation)

Not a redox reaction – fermentation reaction

Occurs only in highly anaerobic conditions

 $4C_6H_6 + 18H_2O \rightarrow 9CO_2 + 15CH_4$

Methanogenesis

Indicators of methanogenesis:

Increase in methane and carbon dioxide Reduction in hydrocarbon concentration Very low or no dissolved oxygen Presence of methanogenic bacteria

Carbon dioxide neutralization

All hydrocarbon degradation processes create CO₂

 $\begin{array}{lll} \text{CO}_2 + \text{H}_2\text{O} \rightarrow \text{H}_2\text{CO}_3 & (\text{carbonic acid}) \\ \text{H}_2\text{CO}_3 + \text{CaCO}_3 \rightarrow \text{Ca}^{2+} + 2\text{HCO}^{3-} \end{array}$

CO₂ neutralization increases alkalinity 8 mg alk produced per mg benzene degraded

Ground water:

- Total hydrocarbons confirm HC decrease
- Aromatic hydrocarbons confirm BTEX decrease
- Oxygen confirm utilization, redox state
- Nitrate confirm utilization
- Iron(II) confirm production
- Sulfate confirm utilization
- Methane confirm methanogenesis

Ground water:

- Alkalinity confirm CO₂ production and neutralization
- Oxidation-reduction potential confirm geochemical environment
- pH, temperature, conductivity, chloride confirmation of single ground-water system

Biological:

- Field dehydrogenase test confirm presence of aerobic bacteria
- Volatile fatty acids biodegradation byproduct of complex organic compounds
- Microcosm studies confirm biodegradation is occurring

Average Relative Contribution of BTEX Biodegradation Processes in Site Ground Water at 42 sites



Source: http://www.afcee.brooks.af.mil/er/ert/download/natattenfuels.ppt

Average Relative Contribution of BTEX Biodegradation Processes in Site Ground Water at 42 sites

(Excluding five sites with >200 mg/L Sulfate Reduction Capacity)



Treatability Study Results (continued)

BTEX assimilative capacity averaged 64 mg/L Field-scale biodegradation half-lives: Range = 9 days to 9.5 years Mean = 1 year.



BIOSCREEN



Anaerobic transformation of carbon tetrachloride

 $\begin{array}{c} \mathsf{CCI}_4 \to \mathsf{CHCI}_3 \to \mathsf{CH}_2\mathsf{CI}_2 \to \mathsf{CH}_3\mathsf{CI} \\ \downarrow \\ \\ \mathsf{CO}_2 + \mathsf{H}_2\mathsf{O} + \mathsf{CI}^- \end{array}$

Redox conditions: Denitrification $CCI_4 \rightarrow CHCI_3$ Sulfate reduction $CCI_4 \rightarrow CO_2 + H_2O + CI^-$

Abiotic degradation of TCA

 $CH_3CCI_3 \rightarrow CH_2CCI_2 + CH_3COOH \rightarrow CO_2 + H_2O + CI^-$ 1,1,1-TCA \rightarrow 1,1-DCE + Acetic acid \rightarrow mineralization

Redox conditions: All redox conditions

Anaerobic dechlorination of TCA

 $\label{eq:charge} \begin{array}{l} \mathsf{CH}_3\mathsf{CCI}_3 \to \mathsf{CH}_3\mathsf{CHCI}_2 \to \mathsf{CH}_3\mathsf{CH}_2\mathsf{CI} \to \mathsf{CO}_2 + \mathsf{H}_2\mathsf{O} + \mathsf{CI}^- \\ \texttt{1,1,1-TCA} \to \texttt{1,1-DCA} \to \texttt{Chloroethane} \to \texttt{mineralization} \end{array}$

Redox conditions:

- Sulfate reduction 1,1,1-TCA $\rightarrow 1,1$ -DCA
- Methanogenesis 1,1,1-TCA \rightarrow mineralization

Anaerobic degradation of PCE & TCE

 $\begin{array}{c} \mathsf{CCI}_2 \texttt{=} \mathsf{CCI}_2 \to \mathsf{CHCI}\texttt{=} \mathsf{CCI}_2 \to \mathsf{CHCI}\texttt{=} \mathsf{CHCI} \to \\ \mathsf{CH}_2 \texttt{=} \mathsf{CHCI} \to \mathsf{CH}_2 \mathsf{CH}_2 \to \mathsf{CH}_3 \mathsf{CH}_3 \end{array}$

 $\begin{array}{rccc} \mathsf{PCE} & \to & \mathsf{TCE} & \to & \mathsf{cis-1,2-DCE} \to \\ & & \mathsf{vinyl\ chloride} \to \mathsf{ethene} \to \mathsf{ethane?} \end{array}$

Redox conditions:

Sulfate reduction $PCE \rightarrow DCE$, $TCE \rightarrow DCE$ Methanogenesis $PCE \rightarrow ethene$, $TCE \rightarrow ethene$

Ground water:

- Same as for hydrocarbons, plus:
- Methane to confirm methanogenesis
- Chlorinated VOCs materials degraded and degradation products
 - Distinguish cis-1,2-DCE and trans-1,2-DCE
- Degradation byproducts: CO₂, ethane, ethene, chloride

Ground water:

Dissolved hydrogen – distinguishes strength of dechlorination and redox state:

Electron acceptor	H ₂ concentration (ng/L)
Denitrification	< 0.1
Iron reduction	0.2 to 0.8
Sulfate reduction	1 to 4
Reductive dechlorination	> 1
Methanogenesis	5 to 20

Extent of Chlorinated Solvents and BTEX



BTEX and Electron Acceptors



BTEX and Metabolic Byproducts



Chlorinated Solvents and Byproducts



Plot of TCE, DCE, VC, and Ethene versus Distance Downgradient



Plume Classification















BIOCHLOR



Diffusive disappearance

Parker, B. L., R. W. Gillham and J. A. Cherry, 1994. "Diffusive Disappearance of Immiscible-Phase Organic Liquids in Fractured Geologic Media." *Ground Water*, Vol. 32, No. 5, Pp. 805-820. September/October 1994. See Figure 1 of:

Parker, B. L., R. W. Gillham and J. A. Cherry, 1994. "Diffusive Disappearance of Immiscible-Phase Organic Liquids in Fractured Geologic Media." *Ground Water*, Vol. 32, No. 5, Pp. 805-820. September/October 1994.



$$J_{D} = -nD_{e}\frac{\partial C}{\partial x}$$
$$\frac{\partial C}{\partial t} = \frac{D_{e}}{R}\frac{\partial^{2} C}{\partial x^{2}}$$

Assume
$$C(x = 0, t) = S_w$$

and $C(x, t = 0) = 0$

$$J_{D}(0,t) = -nS_{W} \left(\frac{RD_{e}}{\pi t}\right)^{1/2}$$

Integrate over time to get mass leaving fracture :

$$M_{t} = nS_{w} \frac{4}{\sqrt{\pi}} (RD_{e}t)^{1/2}$$



Image adapted from: Parker, B. L., R. W. Gillhamand J. A. Cherry, 1994. "Diffusive Disappearance of Immiscible-Phase Organic Liquids in Fractured Geologic Media." *Ground Water*, Vol. 32, No. 5. Pp. 805-820. September/October 1994.

Geologic media parameters for three type-fractured geologic media used for comparison of TCE mass loss rates due to diffusion and sorption to matrix solids presented in figure: type-clay, type-shale/sandstone, and type-granite.



Image adapted from: Parker, B. L., R. W. Gillhamand J. A. Cherry, 1994. "Diffusive Disappearance of Immiscible-Phase Organic Liquids in Fractured Geologic Media." *Ground Water*, Vol. 32, No. 5. Pp. 805-820. September/October 1994.